

SPATIAL BIOECONOMIC DYNAMICS OF MARINE FISHERIES

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Keywords: Spatial bioeconomic analysis, source-sink, metapopulations, marine protected areas, spatial fisheries management.

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Summary

Spatial management of fisheries requires understanding the spatial behaviour of species with the corresponding abundance heterogeneity in space and time and the ecological interdependencies within an ecosystem framework. It also involves proper understanding of fisher behaviour driving the spatial allocation of fishing intensity. The recognition of the implications of dynamic pool assumptions in over estimating stock abundance, is discussed together with spatial modelling efforts aimed at relaxing this unrealistic assumption for sedentary species and many low mobility demersal resources. Progress aimed at considering management implications of metapopulations and source-sink configurations present in many marines populations, is also discussed. The establishment of marine reserves as a strategy for enhancing the conservation of marine resources is examined with respect to the implications of their size and location when considering metapopulations and source-sink configurations. Recent reports of the benefits and costs of Marine Protected Areas are also summarised.

1. Introduction

Understanding the spatial dynamics of marine fisheries allows proper management of fish resources that distribute heterogeneously in space and time. This consideration is relevant for most marine species, but critical for sedentary marine resources. Dynamic pool fishery models developed in the 1950's, have been criticized as unrepresentative of

events in a fishery in that spatial realism is sacrificed. Authors of these models, nonetheless laid the foundations for incorporating spatial considerations into population dynamics, but in the 1950's were unable to follow up their ideas due to the inadequacy of computational facilities. In general, these models are based on dynamic pool assumptions, which establish that: (i) the resource is homogeneously distributed in space; (ii) ages are perfectly mixed; and either (iii) fishing effort is applied uniformly over the range of resource distribution, or (iv) after fishing effort has been applied, the resource is able to redistribute itself according to (i) and (ii). Studies typically confirm that local habitats are unequal in quality and holding capacity throughout the stock range. Not surprisingly, for sedentary resources, models based on dynamic pool assumptions are inadequate and result in serious model error. The spatial distribution of these resources is patchy, in terms of size, density and age structure. As a result, the allocation of fishing effort is spatially heterogeneous. The principal consequence of this spatial heterogeneity is that under dynamic pool assumptions, the productive potential of the stock is overestimated, increasing the risk of over-exploitation and collapse of the fishery.

Spatial components have subsequently been introduced into population models for a variety of motives: to explore sequential fisheries, to examine spatial allocation strategies, and to consider the implications of optimal foraging theory. Spatial modelling exercises introducing elements of spatial realism, both in the distribution pattern of the resource and the fishing strategy, usually result in the management implications for age-structured models less optimistic.

2. Models of exploited populations incorporating spatial structure

Dynamic pool models suppose that each individual has an equal probability of mating and spawning, and this has been referred to as panmixia. However, more recent studies of marine fish populations using trace elements, parasites and genotypes have begun to discover genetic differentiation within demersal marine resources of continental shelves, suggesting that metapopulations are fairly common, especially for sedentary or territorial species. Hence, the effective reproductive size of a population may be much smaller than its total population size and reproductive age groups and spawning sites may not be equally successful in their reproductive activities throughout the species range. One common variant is where a progressive isolation of source areas occurs with distance, leading eventually to the separation of local genotypes.

To deal with this complexity an early spatial model, YAREA, explored harvest strategies for spatially-differentiated resources and fishing strategies, and was developed further under a variety of assumptions to address both resource and bioeconomic considerations. The authors developed a suite of models for exploring the implication of port location and distance from fishing grounds and a bioeconomic simulation of age-structured spatial populations, and an explicit consideration of spatial considerations was later extended to optimizing rotating harvest management strategies. General models of how spatial considerations affect stock and fleet dynamics were also developed. Some studies have reviewed how geographical considerations affect exploitation of marine populations. More recently, scientists modelled the heterogeneity

in resource and effort distribution using the negative binomial distribution in an environment of risk and uncertainty.

3. Heterogeneous recruitment density in space and time

The heterogeneous distribution of recruits over space can be modelled by multiplying the estimated number of recruits produced by the spawning stock (SS_t) over time $R_t = f(SS_t)$ (e.g. estimated using Ricker, Beverton-Holt or stochastic recruitment functions) by a probability density function that distribute them over space.

For instance, a general Beverton-Holt stock-recruitment function can be multiplied by an appropriate probability distribution (i.e. negative binomial distribution, that allows for patches of zero recruitment) to generate a heterogeneous recruitment density over time, as expressed by Equation (1) as follows:

$$R_t = \frac{B_t \cdot \alpha}{\beta + B_t} \cdot P(d) \quad (1)$$

Where B_t is the total spawning biomass over time, α is maximum annual recruitment, and β is total spawning biomass for $\alpha/2$. $P(d)$ is the negative binomial probability density function. An important assumption of Equation (1) is that recruitment depends on the total spawning biomass rather than the locus-specific spawning biomass. This assumption is likely to be valid when the life cycle of species involves indirect development, that is, when juveniles do not emerge directly from the egg but rather as a result of metamorphosis of larvae that can recruit in a different locus from the one inhabited by the parental stock.

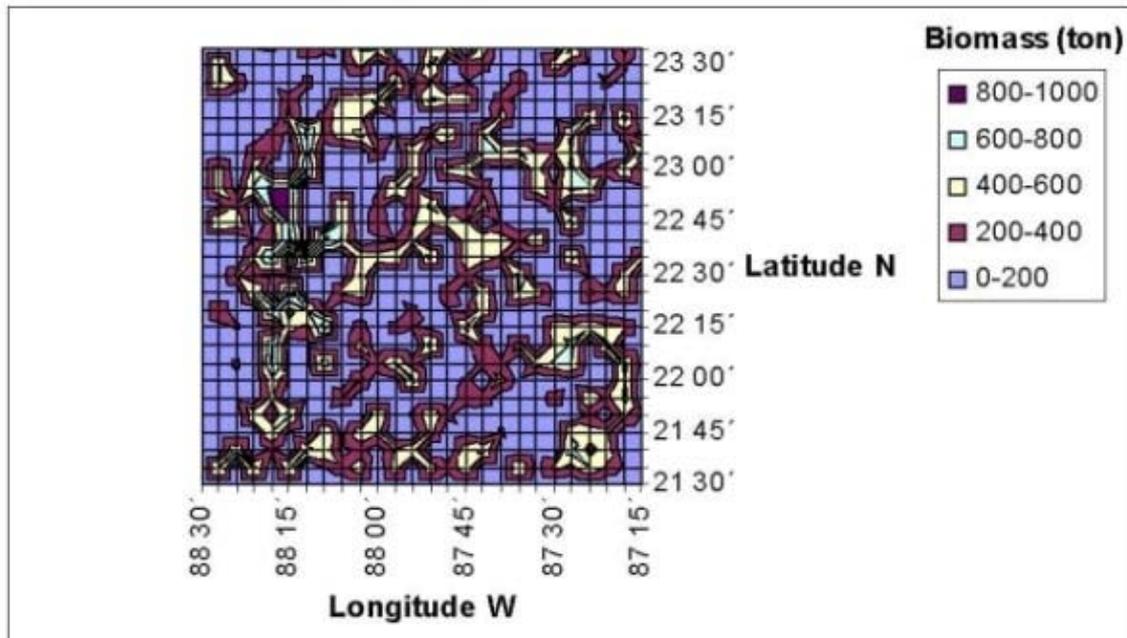


Figure 1. Spatial distribution of biomass using a negative binomial function for recruits settlement over space.

The negative binomial distribution used to represent spatial heterogeneity in recruitment densities is estimated as follows:

$$P(d) = \left(1 + \frac{\mu}{\varepsilon}\right)^{-\gamma} \cdot \frac{(\varepsilon + d - 1)!}{d!(\varepsilon - 1)!} \cdot \left(\frac{\mu}{\mu + \varepsilon}\right)^d \quad (2)$$

Where d represents recruitment density, ε is the family member of the negative binomial, and μ is the mean recruitment density. The basic assumption is that the stock can be subdivided into ‘loci’, each assuming different resource densities. A ‘locus’ here is effectively the smallest geographical unit considered. Each locus would contain several age classes, all of which would have different densities. ‘Locus’ can be considered synonymous with ‘cell’, ‘quadrat’ or ‘pixel’ in common usage, and can be assigned a specific geographical position or latitude/longitude (Figure 1).

4. Spatial allocation of effort

Modelling the short-run spatial dynamics of marine fisheries allows for better understanding the inter-temporal allocation behaviour of fishing effort and thus to develop adequate management strategies.

Some spatial allocation strategies documented in the literature include the following:

- Proportional allocation according to the spatial abundance of the resource.
- Sequential allocation to those patches of greatest abundance
- Random search
- Free distribution of allocation of fishing intensity
- Proportional allocation to:
 - a. The quasi-rent of the variable costs (including transfer costs resulting travelling from port to alternative fishing grounds).
 - b. The friction of distance, i.e. non-monetary costs associated to vessel distance travelled to fishing grounds
 - c. The probability of finding the target species in profitable levels.

This last spatial allocation strategy, involving a short-run effort allocation decision, is expressed in equation (3) as follows,:

$$SAE_{kht+1} = \frac{P_k \cdot \text{quasi}\pi_{kht} / D_{hk}^{\phi}}{\sum_k P_k \cdot \text{quasi}\pi_{kht} / D_{hk}^{\phi}}$$

Where, P_k is the probability of finding the target species in profitable levels in alternative fishing grounds k . The *quasi* rent of the variable costs (rent obtained after paying for the variable costs of fishing) received by the average vessel leaving port h by fishing in site k in time t is expressed as *quasi* π_{hkt} . Finally, D_{hk} represents the distance from port h to fishing site k , and ϕ the friction of distance parameter that accounts for the non-monetary costs associated to distance travelled to fishing site.

The properties of this simple spatial allocation model (SAE) are summarized in Table 1 for alternative fishery cases.

Fishery Case	Monetary costs associated to distance D	Friction of distance ¹ (ϕ)	Probability of finding the resource (P)	Expected spatial distribution of effort
Case 1	$D_1 = D_2 = \dots = D_k$ $\theta = 0$	$\phi = 0$	$P_1 = P_2 = \dots = P_k$	Proportional to resource abundance
Case 2	$D_1 \neq D_2 \neq \dots \neq D_k$ $\theta > 0$	$\phi = 0$	$P_1 = P_2 = \dots = P_k$	Proportional to the quasi-rent of the variable costs
Case 3	$D_1 \neq D_2 \neq \dots \neq D_k$ $\theta > 0$	$\phi > 0$	$P_1 = P_2 = \dots = P_k$	Proportional to the quasi-rent of the variable costs and $1/D^\phi$
Case 4	$D_1 \neq D_2 \neq \dots \neq D_k$ $\theta > 0$	$\phi > 0$	$P_1 \neq P_2 \neq \dots \neq P_k$	Proportional to the quasi-rent of the variable costs, $1/D^\phi$ and P_k

¹ Non-monetary cost associated to distance.

Table 1. Alternative strategies of spatial allocation of effort.

Considering the SAE_{kht} function at least four types of distributions that account for spatial variations in fishing intensity can be described:

4.1 Case 1: Small-scale littoral fisheries.

For intertidal and sandy beaches bivalve fisheries where distance from port to alternative fishing areas becomes irrelevant in terms of transfer costs from ports to alternative fishing areas ($\theta = 0$), the corresponding friction of distance is also zero ($\phi = 0$), and the probability of find the target species at profitable levels in alternative sites (P_k) is not significantly different, then the resulting SAE_{kht} distribution is *proportional to the spatial variations in stock abundance*.

4.2 Case 2: Small scale fisheries in bays, coastal lagoons and estuaries.

Where transfer distances from port to alternative fishing grounds are relevant ($\theta > 0$), but non-monetary costs are negligible (friction of distance, $\phi = 0$) because of easy fishing operation and navigability in naturally protected (from wind and wave action) fishing grounds, and again the probability of find the target species at profitable levels in alternative fishing sites is again not significantly different, then the resulting SAE_{kht} distribution is *proportional to spatial variations of the quasi-rent of the variable costs*.

4.3 Case 3: Fisheries in exposed coastal zones.

When ($\theta > 0$), the friction of distance is substantial ($\phi > 0$) and $P_1 = P_2 = \dots = P_k$, the SAE_{kht} distribution is proportional to the quasi-rent of the variable costs, and inversely related to the friction of distance from port alternative fishing grounds.

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Bibliography

Amezcuca, A.B., Holyoak, M. 2000. Empirical evidence for predator-prey source-sink dynamics. *Ecology*, 81, pp 3087-3098. [This study tested the prediction suggesting that a source-sink structure for a prey species can promote the persistence of an otherwise nonpersisting predator-prey interaction. Evidence suggests that continuous prey immigration into predator-prey bottles from extinction-invulnerable prey-only bottles may have weakened the coupling between predator and prey dynamics and contributed to the increase in persistence. In showing that source-sink dynamics enhanced predator-prey persistence, results support conclusions of metapopulation theory that point to the importance of immigration between spatially discrete populations.]

Anderson, L.G. 2002a. A Bioeconomic Analysis of Marine Reserves. *Natural Resource Modeling*, 15(3), pp 311-334.[extended the analysis of protected areas using sustainable catch and revenue curves to illustrate how marine reserves influence the proportion of stock available for harvest, and compared fishery performance with or without marine reserves].

Anderson, L.G. 2002b. A comparison of the utilization of stocks with patchy distribution and migration under open access and marine reserves: an extended analysis. *Marine Resource Economics*, 17 pp 269-289. [The author developed a discrete bioeconomic model for two patches with source-sink configuration and developed the corresponding analytical solution for the open access bioeconomic equilibrium for both, density dependent migration and source – sink migration].

Caddy, J.F. 1975. Spatial model for an exploited shellfish population, and its application to the Georges Bank scallop fishery. *J. Fish. Res. Bd. Canada*, 32, pp 1305-1328. [The author built the well known YRAREA that model geographically contiguous unit areas, and the realism of age structure. It generate a virgin stock by stochastic patch distribution of recruitment using the bi-variate normal distribution to distribute individuals over space once the patch center has been randomly selected. Spatially distributed fishing intensity is represented as a function of spatially heterogeneous catch per unit of effort].

Caddy, J.F. Seijo, J.C., 1998. Application of a spatial model to explore rotating harvest strategies for sedentary species. *Canadian Special Publications on Fisheries and Aquatic Sciences*, 125, pp 359-265. [The paper explores the effect of alternative rotating harvest strategies for harvesting of sedentary species with different life cycles. A spatial age structured bioeconomic model was built to explore alternative sizes of closed areas and the optimum rotating period.]

Caddy, J.F. Carocci, F., 1999. The spatial allocation of fishing intensity by port-based inshore fleets: a GIS application. *ICES J. Mar. Sci.* 56, p 388-403. [This paper illustrates some practical Geographical Information System (GIS) applications for aiding fishery managers and coastal area planners in analyzing the likely interactions of ports, inshore fleets, and local non-migratory inshore stocks, and in providing a flexible modelling framework for decision making on fishery development and zoning issues. The classic geographical "friction of distance" approach to generating fields of action around home ports of inshore fleets which largely make day trips to their adjacent fishing grounds, is compared with a more flexible empirical "Gaussian Effort Allocation" (GEAM) modelling approach where peak effort may occur at different distances from port. The latter approach is considered more appropriate in describing resource depletion with distance. The GEAM model is also suggested as an aid to deciding on the location of marine parks or fishery closure areas.]

Charles, A.T., Reed, W.J. 1985. A bioeconomic analysis of sequential fisheries: competition, coexistence, and optimal harvest allocation between inshore and offshore fleets. *Can. J. Fish. Aquat. Sci.*, 42, pp 952-

962. [A bioeconomic model is developed to determine optimal harvest allocation between "offshore" and "inshore" fleets exploiting a single fish stock in sequential fisheries. The socially optimal policy for maximizing total discounted rent is determined in terms of optimal escapement levels in each fishery. Whether exclusion or coexistence of the two fleets occurs under open access and under optimal management is found to depend primarily on inshore/offshore price and cost ratios, together with biological parameters related to the age structure of the fish stock. The authors discuss how fishery regulations, such as separate landings taxes imposed on each fleet, can be used to jointly optimize open-access exploitation in sequential fisheries.]

Clark, C.W. 2006. *The worldwide Crisis in Fisheries: Economic Models and Human Behavior*. Cambridge University Press, 263 pp. [Perspective of simple bioeconomical modelling, regulation of fishing effort, overcapacity, subsidies, etcetera. Dynamic bioeconomical models, investment and overcapacity, fisheries management, risk assessment and risk management, and a few case studies.]

Gillis, D.M., R.M. Peterman, A.V. Tyler. 1993. Movement dynamics in a fishery: application of the ideal free distribution to spatial allocation of effort. *Can. J. Fish. Aquat. Sci.* 50:323-333. [The authors used data on the Hecate Strait, British Columbia, Canada, trawl fishery to test hypotheses about spatial allocation of effort and interaction among fishing vessels. The ideal free distribution of Fretwell and Lucas was the foundation for deriving these tests. The paper found evidence for competition among vessels, although we could not distinguish whether the mechanism was interference or exploitation competition. CPUE was generally equalized among the areas fished, as predicted by the ideal free distribution, because of movement of boats among areas. Thus, area-specific CPUE would not be a reliable index of relative abundance of fish in different areas; relative fishing effort may be better.]

Hannesson, R. 1998. Marine reserves: What do they accomplish?. *Marine Resource Economics*, 13(3), pp 159-170. [This paper investigates what will happen to fishing outside the marine reserve and to the stock size in the entire area as a result of establishing a marine reserve. Three regimes are compared: (i) open access to the entire area, (ii) open access to the area outside the marine reserve, and (iii) optimum fishing in the entire area. Two models are used: (i) a continuous-time model, and (ii) a discrete-time model, both using the logistic growth equation. Both models are deterministic equilibrium models. The conservation effect of a marine reserve is shown to be critically dependent on the size of the marine reserve and the migration rate of fish. A marine reserve will increase fishing costs and overcapitalization in the fishing industry, to the extent that it has any conservation effect on the stock, and in a seasonal fishery it will shorten the fishing season.]

Hannesson, R. 2002. The economics of marine reserves. *Nat. Res. Mod.* 15 (3), pp 273-290. [The effects of marine reserves with open access elsewhere are analyzed, using a logistic model for a population with a patchy distribution. It is assumed that a marine reserve is established for the territory of one of two sub-populations which interact through migrations. The total population increases while the total catch declines for the most part. A high rate of migration would, however, dilute the conservation effect. Examining a stochastic variant of the model shows that the variability (sum of squared deviations) of catches may decrease as a result of protecting one of the sub-populations. Even if all rents disappear by assumption, it is possible to identify this as an economic benefit, particularly when the average catch increases.]

Hilborn, R., Walters, C.J. 1987. A general model for simulation of stock and fleet dynamics in spatially heterogeneous fisheries. *Can. J. Fish. Aquat. Sci.* 44, 1366-1369.

Holland D. y R. Brazee. 1996. Marine reserves for fisheries management. *Marine Resource Economics* 11(3), pp 157-171. [The authors develop a dynamic model of marine reserves applicable to inshore fisheries. In contrast to previous models of reserves, the model is fully dynamic and provides information on both equilibrium conditions and the path to equilibrium. A simulation model based on red snapper data from the Gulf of Mexico is presented. The simulation results suggest that marine reserves can sustain or increase yields for moderate to heavily fished fisheries but will probably not improve yields for lightly fished fisheries.]

MacCall, A.D. 1990. *Dynamic Geography of Marine Populations*. U. Wash. Press, Seattle, Washington. [This book presents a theoretical model linking geography, fish abundance, and population growth. It combines mathematical formulations of habitat selection and population density with standard population growth equations to create a geographic description of fish population dynamics. The model is implemented using real-world fishery data on anchovy populations from the California coast. The key to

the model is a well-known ecological concept called "density dependent habitat selection", which is used at the population level to create a "basin model".]

National Research Council. 2001. *Marine Protected Areas: Tools for Sustaining Ocean Ecosystems*. National Academy Press, 272 pp. [Declining yields in many fisheries and decay of treasured marine habitats, such as coral reefs, has heightened interest in establishing a comprehensive system of marine protected areas (MPAs)-areas designated for special protection to enhance the management of marine resources. Marine Protected Areas compares conventional management of marine resources with proposals to augment these management strategies with a system of protected areas. The volume argues that implementation of MPAs should be incremental and adaptive, through the design of areas not only to conserve resources, but also to help us learn how to manage marine species more effectively.]

Pulliam, H.R. 1988. Sources, sinks, and population regulation. *Amer. Nat.* 132, pp 652-661. [Animal and plant populations often occupy a variety of local areas and may experience different local birth and death rates in different areas. When this occurs, reproductive surpluses from productive source habitats may maintain populations in sink habitats, where local reproductive success fails to keep pace with local mortality. For animals with active habitat selection, an equilibrium with both source and sink habitats occupied can be both ecologically and evolutionarily stable. If the surplus population of the source is large and the per capita deficit in the sink is small, only a small fraction of the total population will occur in areas where local reproduction is sufficient to compensate for local mortality. In this sense, the realized niche may be larger than the fundamental niche. Consequently, the particular species assemblage occupying any local study site may consist of a mixture of source and sink populations and may be as much or more influenced by the type and proximity of other habitats as by the resources and other conditions at the site.]

Pezzey J.C.V., Roberts C.M., Urdal, B.T. 2000. A simple bioeconomic model of a marine reserve. *Ecological Economics*. 33, pp 77-91. [The authors model the effect of a no-take reserve in a marine fishery management area, such as on a coral reef. Implicitly, eggs and larvae are mobile but adults are not; and there is open access fishing outside the reserve. A reserve is found to increase equilibrium catch if the prior ratio of stock to carrying capacity is less than a half, and the catch-maximising reserve proportion rises towards a half as this ratio falls towards zero. After initial adjustment, long-run stability is improved by a reserve. They estimate that coral reef reserves could increase world wide annual catches by about a billion dollars]

Roberts, C. M., Sargant H. 2002. Fishery benefits of fully protected marine reserves: Why habitat and behaviour are important. *Natural Resources Modeling*. 15 (4), pp 487-507. [Fully protected marine reserves, areas that are closed to all fishing, have attracted great interest for their potential to benefit fisheries. The authors present a simple model of reserve effects on a migratory fish species. The model incorporates spatial variation in vulnerability to capture and shows that strategically placed reserves can offer benefits in the form of increased spawning stock and catch, especially when fishing intensities are high.]

Sanchirico, J.N., Wilen, J.E. 2001. Dynamics of spatial exploitation: A metapopulation approach., *Natural Resource Modeling* . 14 (3), pp 391-418. [The authors present a bioeconomic model of a harvesting industry operating over a heterogeneous environment comprised of discrete biological populations interconnected by dispersal processes. The model yields a simple, but insightful, framework from which one can investigate factors that contribute to the evolution of resource exploitation patterns over space and time. For example, we find that exploitation patterns are driven by biological and fleet dispersal and biological and economic heterogeneity. Authors conclude that one cannot really understand the biological processes operating in an exploited system without knowing as much about the harvesting system as about the biological system.]

Seijo, J.C., Caddy, J.F., Euan, J. 1994. Space-time dynamics in marine fisheries: a bio-economic software package for sedentary species. *FAO Computerised Information Series, Fisheries*. FAO Rome, 116p + discs. [The authors present a simulation package developed to model the space-time distribution of fishing intensity using alternative approaches. ALLOC is a short-run spatial bioeconomic model that represents the interdependencies of small-scale and industrial fleets from different ports of origin, harvesting a target species over several fishing grounds. CHART models the short and long-run spatial dynamics of sedentary and low mobility demersal resources as a result of interacting biologic, economic

and geographic characteristics. It is an age structured spatial bioeconomic model that estimates distance and transfer costs from different ports of origin to alternative fishing sites. It models seasonality of recruitment and fishing intensity using the distributed delay model. Allocates seasonal effort over space and time using a function considering the quasi-rent of the variables costs obtained from different sites in previous trip, the probability of finding the resource in profitable levels in alternative sites, and the friction of distance to account for the non-monetary costs of fishing.

Seijo, J.C., Defeo, O., Salas, S. 1998. Fisheries Bioeconomics: Theory, modelling and management. *FAO Fish. Tech. Pap.* 368, pp 107. [In Chapter 6 of this book, the dynamic pool assumption of bioeconomic models is relaxed to model the spatial dynamics of marine species and the corresponding distribution of fishing intensity over space and time].

Seijo, J.C., Pérez, E., Caddy, J.F. 2004. A simple approach for dealing with dynamics and uncertainty in fisheries with heterogeneous resource and effort distribution. *Marine and Freshwater Research (CSIRO Publishing)* 55, pp 249-256. [A spatial bioeconomic model based on the negative binomial distribution to represent patches that vary in size, density and age structure over time. The model incorporates decision theory and different levels of risk aversion in resource management to account for the uncertainty associated with alternative spatially disaggregated fishing strategies.]

Smedbol, R.K., McPherson, A., Hansen, M. M., Kenchington, E. 2003. Myths and moderation in marine metapopulations. *Fish and Fisheries*, 3, 20-35. [This paper summarizes the concept of metapopulation dynamics and the associated theoretical assumptions. We call for a stricter definition and use of the term 'metapopulation', critically evaluate the applicability of metapopulation theory to marine population dynamics and its use in the related literature, and consider two published case-studies that investigate metapopulation structuring in specific marine populations. Authors urge scientists to carefully articulate what is meant by the term 'metapopulation' and to use appropriate citations in the primary literature to circumvent the potential for nebulous (and possibly damaging) conclusions in the future.]

Smith, M.D., Wilen J.E. 2004. Marine reserves with endogenous ports: empirical bioeconomics of the California sea urchin fishery. *Marine Resource Economics*. 19 (1), pp 85-112. [This paper adds another layer of behavioral realism to the bioeconomics of marine reserves by endogenizing fisher home port choices with a partial adjustment share model. Estimated with Seemingly Unrelated Regression over monthly data, this approach allows simulation of both short- and long-run behavioral response to changes induced by marine reserve formation. The findings cast further doubt on the notion that marine reserves generate long-run harvest benefits.]

Sumaila, U.R. 1998. Protected marine reserves as fisheries management tools: a bioeconomic analysis. *Fisheries Research* 37, 287-296. [This paper develops a dynamic computational bioeconomic model with the objective of assessing protected marine reserves as fisheries management tools. Two key results emerge from the study. First, establishment of marine reserves are bioeconomically beneficial when net transfer rates for cod are 'reasonably' high and reserve sizes are large: large reserves provide good protection for the stock in the face of the shock, while high transfer rates make the protected fish available for harvesting after the shock has occurred. Further, optimally chosen reserve size when net transfer rates are high, also mitigates against biological losses. Second, when net transfer rates are low, the establishment of marine reserves does not mitigate against losses in the discounted economic rent, while they tend to be efficient in mitigating against biological losses.]

Wilen, J. 2004. Spatial Management of Fisheries. *Marine Resource Economics*, 19(1):7-20. [This paper discusses recent advancements in scientific understanding about the spatial distribution of abundance in the ocean and the processes that determine abundance. A new spatial management paradigm is envisioned whereby electronic vessel and gear monitoring allows management of effort at fine temporal and spatial scales. The research challenges of this new vision of future management are then discussed, focusing on understanding spatial behavior of fishermen, developing integrated spatial bioeconomic management models, and exploring alternative management instruments for regulating the spatial distribution of harvesting.]

Biographical Sketch

Juan Carlos Seijo is Professor of the School of Natural Resources, Marist University of Merida where he was University Rector from 1996 to 2004. He received his MSc. (1979) and Ph.D. (1986) degrees in

Resource Economics from Michigan State University. He has taught for 20 years graduate and undergraduate courses in Fisheries and Aquaculture Bioeconomics. His academic work has been published in scientific journals that include *Marine Resource Economics*, *Journal of Aquaculture Economics and Management*, *Fisheries Research*, *Philosophical Transactions of the Royal Society*, among others. He is author of two books in his field of specialization. He has taught specialized courses in bioeconomics organized by FAO and UNESCO in Chile, Uruguay, Peru, Colombia, Venezuela, Panama, Guatemala, Cuba, and Trinidad, and has participated in Expert Consultations invited by FAO in Lysekil, Sweden (1995), Australia (1998), Rome (2000), Mauricio (2003), and Cambodia (2004). He has been guest and visiting professor in the Ocean University of Taiwan (Keelung), Center for Marine Studies of the University of Delaware, and the Institute of Aquaculture of the University of Stirling. He is currently Chairman of the Scientific Advisory Group of WECAFC (West Central Atlantic Fisheries Commission), and Board member of the North American Association of Fisheries Economists (NAAFE).